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4 **Guideline on Risk characterisation and assessment of**
5 **Maximum Residue Limits (MRL) for biocides**
6 **Draft**

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30 **Executive summary**

31 Where it is considered that residues of pharmacologically active substances¹ in biocidal products used
32 in animal husbandry might have the potential to lead to consumer health concerns a consumer safety
33 evaluation must be undertaken with, where appropriate, the derivation of maximum residue limits
34 (MRLs). This document briefly introduces the process by which a decision is taken on whether an MRL
35 evaluation is needed and details the approach taken for the MRL evaluation.

36 A step-wise procedure is used to determine whether an MRL assessment is required for a biocidal
37 substance used in animal husbandry². The procedure uses a threshold concept for external exposure of
38 food producing animals to identify those substances for which an MRL evaluation is needed and allows
39 refinement of the external exposure estimate based on relevant data. If the estimated external
40 exposure of a food producing animal to the pharmacologically active substance and/or its toxic
41 degradation products and/or any substance of concern contained in the biocidal product exceeds the
42 trigger value (of 4 µg/kg bw), this is interpreted as indicating a possible consumer risk of residues and
43 triggers a request for a formal MRL procedure. If, on the other hand, the external exposure is below
44 the trigger value then, in most cases, there will be no need for an MRL evaluation. However, all hazard
45 endpoints need to be carefully considered in making this decision, and if the pharmacologically active
46 substance presents a particular concern, then the trigger value of 4 µg/kg bw/day for external
47 exposure of the animal is not considered sufficiently protective and consequently an MRL evaluation
48 would need to be undertaken. It should be noted that for substances considered to induce non-
49 threshold toxicity effects (either directly or indirectly via metabolites) such as genotoxicity it will
50 usually not be possible to establish an ADI or MRLs.

51 In those cases where it is determined that an MRL evaluation is required, the responsibility for
52 undertaking the MRL evaluation falls to the European Medicines Agency's Committee for Medicinal
53 Products for Veterinary Use (CVMP). The CVMP also uses a stepwise procedure in its evaluation. A final
54 ADI or equivalent health based reference value covering all relevant endpoints is required for this
55 evaluation, which compares the estimated worst case consumer exposure to the ADI. Where
56 appropriate data are available it may be possible to use these to refine the initial estimate of consumer
57 exposure. If it is concluded that exposure will be consistently below the ADI without the need for
58 exposure reduction measures, and in the absence of particular risk management concerns then the
59 CVMP may recommend that there is no need to establish specific MRLs for the substance. If, on the
60 other hand, exposure reduction measures are needed in order to ensure that consumer exposure
61 remains below the ADI, then specific MRL values may be recommended.

62 The stepwise approach aims to minimise the number of cases in which a full set of residue data will be
63 required. The level of data required will mainly depend on the type and quantity of the potential
64 residues and their relation to the established exposure limit (i.e., ADI).

65 **1. Introduction (background)**

66 European legislation specifies that biocidal products containing active substances that, as a result of
67 their use, may lead to residues in food shall only be authorised if these residues do not have
68 unacceptable effects on human health and that, where appropriate, an ADI and MRL should be

¹ Regulation No. 470/2009 uses the term 'residues of pharmacologically active substance', which is defined to encompass both residues of active substances and residues of excipients. Directive 98/8/EC uses the terms 'active substance' and 'substance of concern' and so distinguishes between the active substance and other product components. This guideline was developed with a view to facilitating the implementation Regulation No. 470/2009 and consequently the terminology used in this guideline is taken from that regulation. However, for the purposes of this guideline, the term 'pharmacologically active substance' is considered to encompass both the 'active substance' and 'other substances of concern'

² European Commission Draft Guidance on Estimating Livestock Exposure to Active Substances used in Biocidal Products.

69 established. The legislation further states that the European Medicines Agency is the body responsible
70 for performing MRL evaluations for pharmacologically active substances used in biocidal products for
71 use in animal husbandry.

72 The purpose of this paper is to present the approach taken in the MRL evaluation of pharmacologically
73 active substances included in biocidal products for use in animal husbandry and to provide guidance on
74 the type of data required in relation to the dietary risk assessment and MRL evaluation.

75 **2. Scope**

76 Biocidal substances are used in many different situations and residues of biocidal substances may
77 potentially enter the food chain as a result of a number of these uses (including exposure of plants to
78 biocides, exposure of food producing animals to biocides and contamination of food commodities with
79 biocides). The European Medicines Agency is responsible for performing MRL evaluations only for
80 pharmacologically active substances used in biocidal products used in animal husbandry.

81 For the purposes of this guideline, biocidal products used in animal husbandry are considered to be
82 biocidal products used for the purposes of caring for and rearing food producing animals, and to which
83 food producing animals are exposed during some stage of their lifetime. The detailed evaluation of
84 consumer exposure to residues of biocidal substances that occur in food commodities as a result of the
85 use of biocidal products after the end of the animal's life is therefore not considered to be a task for
86 which the European Medicines Agency has responsibility. Similarly, the European Medicines Agency is
87 not considered to be responsible for the detailed evaluation of consumer exposure to residues of
88 biocidal substances that occur as a result of the exposure of milk and eggs to biocidal products after
89 these food commodities have left the animal's body. Nevertheless, when evaluating consumer
90 exposure and establishing MRLs, it is appropriate that any consumer exposure to the substance that
91 may occur as a result of uses of the substance in products other than biocidal products for use in
92 animal husbandry, e.g., use in veterinary drugs, plant protection products or feed additives, is taken
93 into account.

94 **3. Legal basis**

95 Article 5(1)(b)(iii) of Directive 98/8/EC of the European Parliament and of the Council concerning the
96 placing of biocidal products on the market specifies that for biocidal products/active substances that,
97 as a result of their use, may lead to residues in food, Member States shall ensure that products are
98 only authorised if these residues have no adverse effects on human health.

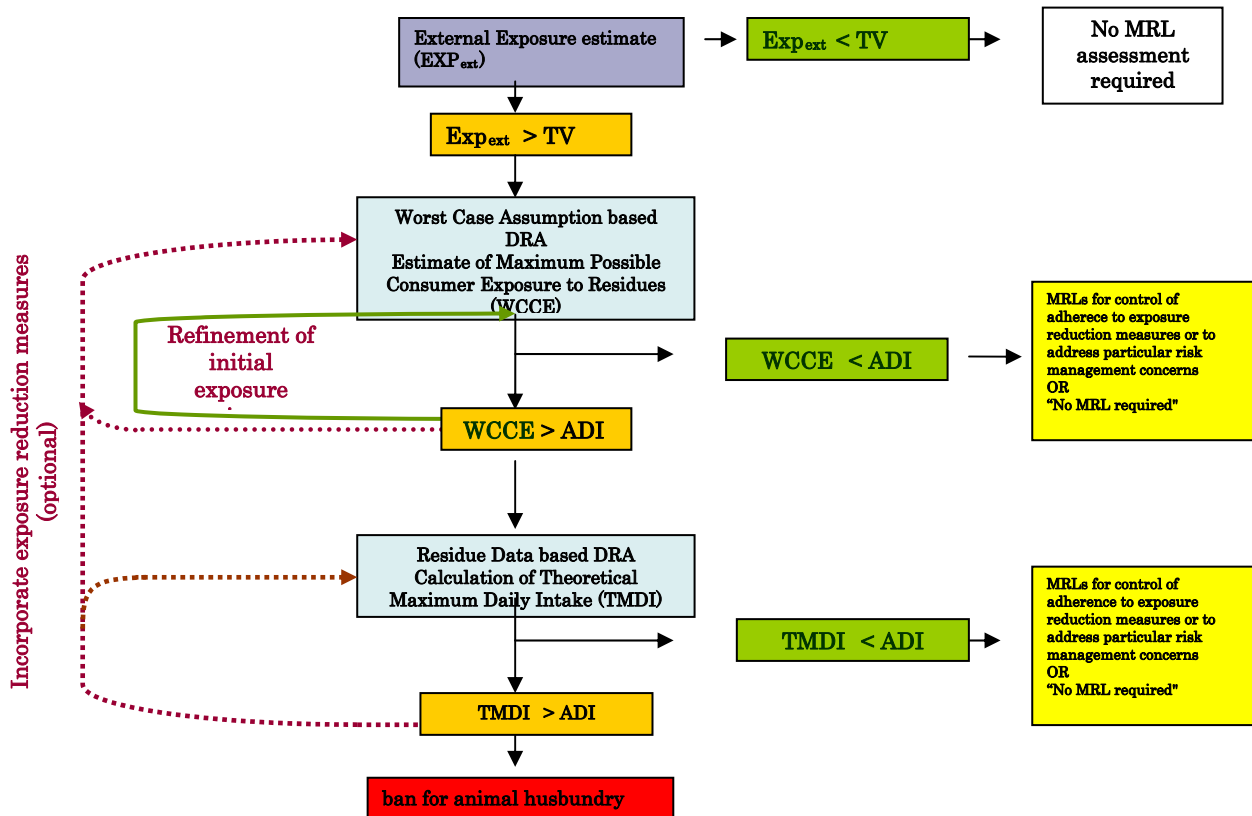
99 Article 10 of Regulation (EC) No 470/2009 of the European Parliament and of the Council provides for
100 the setting of Maximum Residue Limits (MRL) for pharmacologically active substances used in biocidal
101 products used in animal husbandry and specifies that the European Medicines Agency is responsible for
102 recommending MRLs for these substances.

103 **4. Stepwise approach to risk characterisation**

104 **4.1. Decision tree summarising the overall approach**

105 The figure below summarises the overall stepwise approach without specifying which regulatory bodies
 106 are responsible for the different stages (ie the national Competent Authority or the European Medicines
 107 Agency).

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- Exp_{ext}** = **External exposure of the animal**
- TV** = **Trigger Value (4 µg/kg/day)**
- DRA** = **Dietary Risk Assessment**
- WCCE** = **Worst Case Consumer Exposure**
- TMDI** = **Theoretical Maximum Daily Intake³ (based on maximum residue concentrations combined with the standard food basket)**
- ADI** = **Acceptable daily intake**
- WP** = **Withdrawal period**

120 In general terms the possible outcomes of the evaluation summarised above are:

- 121 • If the external exposure is less than the trigger value, then in general there is no need for an MRL
 122 evaluation and the substance is not entered into Commission Regulation (EU) No. 37/2010

³ According to Volume 8, the intake assessment of residues is based on a Theoretical Maximum Daily Intake (TMDI) approach. The TMDI is the sum of residues present in a standard food basket made up of 300 g muscle, 100 g liver, 50 g fat, 50 g kidney plus 1500 g milk, 100 g eggs and 20 g honey. When calculating the TMDI it is assumed that all food commodities in the standard food basket contain residues at the upper end of the residue distribution (for example, at the 95 % tolerance limit). The risk characterization is based on the TMDI/ADI ratio, both in relation chronic and short-term exposure situations. It is noted that this approach differs from the approach used in dietary risk assessments for plant protection products (PPPs) and so may differ from the approach that will be used in dietary risk assessment of biocide residues in products of plant origin.

123 (however, in cases where there is a particular concern in relation to the toxicity of a substance,
124 then an MRL evaluation may be required even when the external exposure is less than the
125 threshold value – see section 4.1.1 for further detail);

126 • If the trigger value is exceeded but it is concluded that consumer exposure to residues (ie the
127 WCCE or the TMDI) will be less than the ADI at all timepoints after application of the product and
128 without implementation of any exposure reduction measures, and in the absence of particular risk
129 management concerns (for example relating to the potential for misuse), then the CVMP may
130 recommend entry of the substance into Commission Regulation (EU) No. 37/2010 with a 'No MRL
131 required' status;

132 • If it is concluded that exposure reduction measures are required in order to ensure that consumer
133 exposure will remain below the ADI or if there are particular risk management concerns (for
134 example relating to the potential for misuse), then the CVMP may recommend entry of the
135 substance into Regulation (EC) No. 37/2010 with specific MRL values calculated to bring the
136 exposure to residues below the ADI, or alternatively the substance may be banned from use in
137 animal husbandry.

138 It should be noted that the evaluation and eventual establishment of the MRL status for an active
139 substance includes consideration of the intended use of the substance. Consequently, if it is considered
140 that consumer exposure to residues will exceed the ADI it may be possible to incorporate exposure
141 reduction measures (as indicated by the dotted line in the schematic) in order to ensure that the ADI is
142 not exceeded. The nature of any proposed exposure reduction measures should be fully described.
143 Where exposure reduction measures are accepted, MRLs should be derived taking these into account.
144 In this way, compliance with the maximum residue limits will demonstrate compliance with the
145 exposure reduction measures (and so ensure consumer exposure to residues at a level below the ADI).

146 **4.1.1. Evaluation of the external exposure of an animal**

147 The Biocides Technical Meeting has established a trigger value for "external" exposure of an animal of
148 4 µg/kg bw/day, summed over all routes⁴ (oral, dermal and inhalation). In the majority of cases, if
149 external exposure is below this trigger value, then it is concluded that there is no need for an MRL
150 evaluation. If, on the other hand, external exposure exceeds this value, then it is considered that the
151 presence of residues in edible products may represent a consumer safety concern, and consequently a
152 dietary risk and MRL assessment will be initiated. However, use of the trigger value is not considered
153 appropriate in the following cases:

- 154 • for substances that exert non-threshold toxicity effects (either directly or indirectly via
155 metabolites) such as genotoxicity - it will usually not be possible to establish an ADI or MRLs.
- 156 • for substances of particular concern (such as substances with
157 reproductive/developmental/neurotoxic actions or effects on other critical endpoints) the external
158 dose trigger of 4 µg/kg bw/day is not considered to be sufficiently protective, and an MRL
159 evaluation should be undertaken regardless of the external exposure level. Substances with the
160 potential to accumulate (eg, substances with a log Pow of greater than 3) may also represent a
161 particular concern. For the purposes of this evaluation substances for which it is estimated that the
162 ADI will be below 5 µg/kg bw⁵ should be considered to be of particular concern⁶.

⁴ The method by which a figure of 4 µg/kg bw/day was reached is shown in Annex 1.

⁵ The method by which a figure of 5 µg/kg bw was concluded to be an appropriate value for defining substances of particular concern in this context is presented in Annex 2.

163 Further guidance on evaluating external exposure of food producing animals to biocidal substances is
164 provided in the European Commission Draft Guidance on Estimating Livestock Exposure to Active
165 Substances used in Biocidal Products. The remainder of this document is dedicated to describing the
166 process and data requirements for the dietary risk and MRL assessment.

167 **4.1.2. Evaluation of consumer exposure and MRL derivation**

168 The assessment is based on risk characterization of residues in animal derived food that may occur
169 following exposure of the animal to the biocidal substance/product.

170 A valid ADI (or equivalent alternative reference value) derived in line with the principles outlined in
171 Volume 8 of the Rules Governing Medicinal Products in the EU (hereafter referred to as Volume 8) is
172 required for this assessment. As the residues to which the consumer will be exposed may differ from
173 the substance for which the ADI was originally established, the applicability of the ADI will need to be
174 assessed in each case, in particular where in-situ degradation or transformation of the active
175 ingredient may be expected to occur.

176 In a first step, a theoretical exposure estimate for the internal dose received by the animal and the
177 resulting residues in commodities of representative food producing species will be made. This estimate
178 will, as first approximation, use worst case assumptions⁷. The resulting (worst case) consumer
179 exposure (WCCE), determined by combining the estimate of the internal dose received by the animal
180 with the standard food basket, would be compared to the ADI (see footnote 3 for information on the
181 standard foodbasket). If required and where appropriate data are available, refinements to the initial
182 estimate can be made in a second step to obtain a refined (more realistic) WCCE.

183 It should be noted that if an ADI or MRLs already exist following evaluation of the substance in relation
184 to its use in another sector (for example, in plant protection products or in feed additives), these
185 existing values will be scrutinised with a view to establishing whether they are compatible with the
186 data provided in relation to use of the substance in a biocide for use in animal husbandry.

187 **4.1.2.1. Worst Case Consumer Exposure (WCCE), refined WCCE and comparison with the** 188 **ADI**

189 Depending on the circumstances that lead to exposure, different exposure scenarios may need to be
190 addressed: in the simplest case of biocidal products for direct treatment of livestock, the exposure
191 scenario would correspond to the intended dosing regimen. For products/substances leading to indirect
192 exposure through the animals' environment, the exposure estimate should be derived from the residue
193 burden for the maximum possible dose and duration of exposure. The estimates should take into
194 account all possible exposure pathways and should also consider residues of the substance that occur
195 as a result of other uses and dietary sources.

196 As a typical worst case, maximum absorption and retention of the substance over time may be
197 assumed. The assumptions about the relative distribution of the substance between the edible tissues
198 of the food basket should be conservative and scientifically plausible. For substances with a known
199 preferential residue formation in certain body tissues this should be taken into account and complete
200 distribution of residues towards the relevant major target tissue may need to be assumed (for
201 example, in the case of highly lipophilic compounds with accumulation/delayed depletion in body fat or
202 certain metals with accumulation in offal tissues).

⁶ At this stage of the evaluation a formal ADI is unlikely to have been established. However, by considering other available information including the Acceptable Exposure Limit (AEL) it should be possible to conclude on whether or not the ADI is likely to be lower than 5µg/kg.

⁷ European Commission Draft Guidance on Estimating Livestock Exposure to Active Substances used in Biocidal Products

203 The WCCE estimate would also need to consider an upper bound of the residue fraction that might be
204 excreted into milk and eggs, when laying or lactating animals are exposed. Experience shows that
205 excretion of xenobiotics into milk and eggs, while primarily dependant on physicochemical properties,
206 is relatively low for most substances and would only comprise a certain fraction of the total dose. The
207 assumption of transfer of the total dose towards these commodities would, in this case, result in an
208 overestimate of the worst case. However, given the wide variety of possible substances, no fixed
209 general default limit can be given here and the worst-case assumptions should be proposed and
210 justified by the applicant on a case by case basis.

211 If the estimate of the WCCE is lower than the ADI (based on appropriately conservative assumptions
212 and margins to cover uncertainties) at all timepoints after application of the product without
213 implementation of any exposure reduction measures and if there are no particular risk management
214 concerns (for example relating to the potential for misuse), then no further assessment of MRLs for the
215 protection of human health would be necessary. In this case the CVMP may recommend that the
216 substance should be included in Table 1 of the Annex of Commission Regulation (EC) No 37/2010, with
217 an MRL entry of "No MRL required".

218 If, on the other hand, the WCCE is greater than the ADI, and if appropriate data are available,
219 refinements to the initial estimate of the WCCE can be made to obtain a refined (more realistic) WCCE.
220 Refinements of an initial WCCE may be based on available ADME data (in particular the extent of
221 absorption/systemic availability, metabolic rates, excretion half-lives, time to reach steady-state levels
222 etc) and consideration of physicochemical parameters of the substance, or on other scientifically
223 justifiable considerations.

224 Appropriate empirical transfer factors may also be used to estimate the maximum transfer of an
225 external dose to edible tissues and in particular into milk and eggs⁸. Experimental data from analogous
226 substances with comparable physicochemical/ADME properties or surrogate data gathered in vitro may
227 also be useful and acceptable when refining an initial worst-case estimate. Assumptions used to
228 calculate the worst case and/or the refined exposure scenarios should be fully explained and justified
229 and the associated uncertainties should be appropriately discussed. Special caution should be taken
230 when worst case scenarios for potentially accumulating substances such as highly lipophilic compounds
231 or accumulating metals are considered.

232 If the estimate of the refined WCCE is lower than the ADI (based on appropriately conservative
233 assumptions and margins to cover uncertainties) at all timepoints after application of the product
234 without implementation of any exposure reduction measures and if there are no particular risk
235 management concerns (for example relating to the potential for misuse), then the CVMP may conclude
236 that no further assessment of MRLs for the protection of human health is required, in which case it
237 may recommend inclusion of the substance in Table 1 of the Annex of Commission Regulation (EC) No
238 37/2010, with an MRL entry of "No MRL required".

239 On the other hand, if it is concluded that the WCCE would exceed the ADI in the absence of exposure
240 reduction measures, then appropriate measures would need to be specified and the WCCE recalculated
241 taking the anticipated effect of these measures into account. If it is concluded that the WCCE would be
242 brought below the ADI as a result of implementation of exposure reduction measures, then numerical
243 MRL values would be set at levels that correspond to the residue limits that would be expected
244 following application of the exposure reduction measures. Compliance with these MRLs would then
245 demonstrate implementation of the exposure reduction measures and ensure that consumer exposure
246 to residues remains below the ADI.

⁸ for example, see Leeman et al. (2007): Transfer of chemicals from feed to animal products: The use of transfer factors in risk assessment. Food additives and contaminants; 24, 1-13.

247 In those cases where MRL values need to be set in order to be able to verify compliance with exposure
248 reduction measures necessary to bring the WCCE below the ADI, it may be possible to set MRL values
249 based on the WCCE estimate and scientifically justifiable assumptions on the approximate tissue
250 residue distribution. This information may be derived using existing kinetic/metabolic data, for example
251 from related food-producing species or laboratory species, or other appropriate literature/data (e.g.,
252 empirical transfer factors). A similar approach may be used when it is considered necessary to set MRL
253 values as a result of risk management concerns (for example, relating to the potential for misuse).
254 Setting MRL values in the absence of genuine residue data in the target species will require the
255 assessor to be confident that the selected marker residue is appropriate and that the relationship
256 between level of the marker residue in a tissue/food commodity and total residues in that tissue/food
257 commodity can be predicted with reasonable confidence. In practice, this is most likely to be the case
258 for substances that are known not to be extensively metabolised. Any estimate based on surrogate
259 data should be sufficiently conservative to account for inherent uncertainties. In the absence of
260 appropriate information, the setting of MRLs at the lowest possible limits (twice the limit of
261 quantification of the analytical method) could also be considered but such an approach would not
262 reflect tissue residue distribution and may be particularly restrictive.

263 Comparing the WCCE (or refined WCCE) to the ADI and bringing the assessment to a conclusion if
264 WCCE is less than the ADI is generally applicable for substances for which the basic metabolic
265 pathways in the food producing specie(s) are known or can be reliably predicted from ADME data and
266 physico-chemical or structural information (e.g., toxicokinetic data, in vitro data, structure-metabolism
267 relationships etc). The data should allow the assessor to conclude with reasonable certainty that the
268 metabolic patterns in the laboratory species (from which the ADI was derived) and in the food
269 producing species are (qualitatively) comparable and that, therefore, the ADI accommodates the
270 pattern of residues likely to occur in the food producing species.

271 **4.1.2.2. The need for residue data**

272 If consumer intake (i.e. the WCCE and the refined WCCE) is calculated to exceed the ADI and
273 incorporation of exposure reduction measures fails to clearly bring the WCCE below the ADI, a
274 conventional dietary risk assessment based on experimental residue data is needed. In this case
275 standard total (radiolabelled) residue studies are required for the relevant species and food
276 commodities (see below).

277 As for the WCCE estimate, the dietary risk assessment performed using residue data should use the
278 theoretical maximum daily intake (TMDI) approach and the standard food basket for commodities of
279 animal origin.

280 It may be assumed that the TMDI is highest at the shortest possible withdrawal period, i.e. at zero
281 withdrawal time, in particular in exposure scenarios mimicking steady state conditions (in practice this
282 means in tissues sampled at up to/around 12 hours after the last dose, plus milk from the first milking
283 and the first eggs laid). Under sub-steady state conditions (eg, single dosing), however, peak levels
284 may not yet have been reached at time 'zero' in all relevant commodities (for example, in eggs) and
285 this should be reflected in the TMDI estimate (TMDI calculated as sum of food basket residues at peak
286 levels in individual commodities: tissues at $t_{\text{zero}/\text{max}}$ plus milk and eggs at t_{max}).

287 If the data demonstrate that the TMDI is lower than the ADI at time zero (t_{zero}) (and subsequent time
288 points) without implementation of any exposure reduction measures and if there are no particular risk
289 management concerns (for example relating to the potential for misuse), then no further assessment
290 of MRLs for the protection of human health would be necessary. In this case the CVMP may
291 recommend that the substance should be included in Table 1 of the Annex of Commission Regulation
292 (EC) No 37/2010, with an MRL entry of "No MRL required".

293 If, on the other hand, it is concluded that the TMDI would exceed the ADI in the absence of exposure
294 reduction measures, or if there is a potential for misuse leading to a TMDI exceeding the ADI, then
295 appropriate measures would need to be specified and the TMDI recalculated (which may require new
296 residue studies) taking the effect of these measures into account. If it is concluded that the TMDI
297 would be brought below the ADI as a result of implementation of exposure reduction measures, then
298 conventional (numerical) MRL values would be set. Compliance with these MRLs would then
299 demonstrate implementation of the agreed exposure reduction measures and ensure that consumer
300 exposure to residues remains below the ADI.

301 If the TMDI cannot be brought below the ADI by implementation of practicable exposure reduction
302 measures, then the substance may need to be banned from use in biocidal products for use in animal
303 husbandry.

304 **5. Data requirements**

305 **5.1. Safety data**

306 An ADI consistent with the requirements and principles outlined in Volume 8 must be established.⁹

307 **5.2. Residue data**

308 The standard residue study is a total radiolabelled residue study (TRR) or other study providing
309 equivalent information (i.e. total residue information), in accordance with Volume 8. The purpose of
310 the study is to obtain a data based dietary risk assessment (DRA) and estimate of the TMDI.

311 Information obtained in the total residue study is also needed to elaborate the MRL (for further details
312 on establishing MRLs see Volume 8).

313 The general design of the studies should conform to the principles set out in Volume 8 and relevant
314 VICH guidelines (where appropriate). Depending on the biocidal substance under consideration and the
315 conditions of exposure, the design for residues studies with biocidal substances may differ in some
316 aspects from the conventional approach for active substances used in veterinary medicinal products,
317 and should consider the points made below.

318 **5.2.1. Total residue studies**

319 Animals

- 320 • If use of the biocidal product will be restricted to a small number of defined species, then total
321 residue studies should be performed using the relevant species only.
- 322 • If use of the biocidal product is not restricted to named species, then, in line with the principles set
323 out in Volume 8 and relevant VICH guidelines (where appropriate), the total residue studies should
324 be performed with at least a representative major ruminant species, a representative monogastric
325 species, and chickens. Residues should be analysed in tissues, milk and eggs (as appropriate) from
326 these species. In addition, data on fish and honey would be required if relevant.

⁹ For the purposes of undertaking MRL evaluations for substances used in biocidal products for use in animal husbandry the ADI must be established in line with the requirements of Annex V of Council Regulation (EC) No. 2377/90 and further detailed in Volume 8. Biocides Directive 98/8/EC also requires the establishment of an ADI where appropriate and at the time of writing, the toxicity data requirements for biocidal substances as laid down in the directive can be considered equivalent to those required by Annex V of Council Regulation (EC) No. 2377/90. However, for certain substances data on additional endpoints not covered by the requirements of Directive 98/8/EC might be needed (i.e. pharmacology data, microbiological data) in order to establish an ADI in line with Volume 8.

327 • Test animals should be representative of the target population for the product. In studies
328 mimicking indirect animal exposure, default body weights of test animals in studies would be
329 approximately in line with the bodyweights listed in Appendix 1, table 1 of the European
330 Commission Draft Guidance on Estimating Livestock Exposure to Active Substances used in Biocidal
331 Products.

332 Test substance and dosing

333 • The test substance should be representative of the substance to which animals are exposed: it may
334 be the active substance or a derivative thereof or a combination of both (for example, if exposure
335 to metabolites/degradation products is an issue).

336 • For substances/products for direct (intended) treatment of animals (for example, repellents, teat
337 dips), residue studies would be performed using the intended product (or an analogous
338 formulation) and dosing schedule.

339 • For substances leading to indirect (unintended) exposure via the animals' environment or
340 food/drinking water, the dosing regimen would need to simulate the actual exposure conditions as
341 closely as possible: the test substance (or substance(s) of concern) should be administered in a
342 suitable form and vehicle that ensures bioavailability and consistent exposure over the duration of
343 the study. The applicant should fully justify the formulation used.

344 • Dose rates should be at least equivalent to the likely maximum daily exposure of the animals (at
345 least greater than or equal to the 95th percentile of the predicted exposure levels). Higher dose levels
346 may be used to accommodate additional uses and exposure scenarios. The choice and dose level
347 should be justified.

348 • In case of multiple exposure routes, studies would need to be conducted for the quantitatively
349 most relevant route, using the combined maximum dose from all exposure routes. In case of
350 situations involving both direct treatment and indirect animal exposure, data are needed to
351 simulate maximum residues for the combined exposure. The default route of administration for the
352 purpose of residue studies is the oral route (even if, for example, real-life exposure is via
353 inhalation).

354 Duration of treatment/Slaughter times/Sampling

355 • Duration of treatment should be long enough to achieve maximum possible residues in all relevant
356 food commodities. For substances for direct (intended) treatment of animals (for example,
357 repellents and teat dips), the duration of residue studies is the maximum treatment period
358 according to proposed product label instructions. If the treatment period is not long enough to
359 reach steady state, the sampling period and spacing of sampling time points after the end of
360 treatment should be appropriate to include peak levels in all relevant commodities. In case of
361 scenarios mimicking continuous or frequent exposure, the dosing period should allow residues to
362 reach steady state. The minimum time needed to reach steady state may be estimated from
363 appropriate pharmacokinetic parameters. In the absence of suitable pharmacokinetic data, the
364 treatment period of the study should be at least 28 days or until residues plateau in milk and eggs,
365 if they have not done so by 28 days¹⁰. The treatment period of the study should be justified.

366 • It is recommended to include a zero slaughter time point (i.e. slaughter up to around 12 hours
367 post dosing – the slaughter time point should be justified based on the depletion kinetics of the

¹⁰ 28 days is in line with the default recommendation for livestock feeding studies:

See OECD 505 "Residues in Livestock" for guidance on duration of feeding studies " *Once acclimatized, animals should be dosed daily for a minimum of 28 days or until residues plateau in milk or eggs, if they have not done so in 28 days*" <http://lysander.sourceoecd.org/vl=2617880/cl=59/nw=1/rpsv/cw/vhosts/oecdjournals/1607310x/v1n7/contp1-1.htm>

368 substance) if a claim is to be made that a substance does not present residues that are of human
369 health concern in edible tissues and that consequently setting of an MRL is not necessary for the
370 protection of human health. Milk and eggs should be collected throughout the period of the study
371 or at least until peak or plateau levels have been reached.

- 372 • For substances that occur naturally or as ubiquitously present environmental contaminants, it is
373 recommended to take milk or eggs from all animals before treatment in order to determine
374 baseline levels of residues. It is also desirable to determine baseline levels in tissues of control
375 animals.

376 **5.2.2. Marker residue studies**

377 Marker residue studies are required only for substances and in species or commodities for which
378 numerical MRLs are to be established. Where these studies are required they should conform to the
379 guidance provided in Volume 8 and relevant VICH guidelines. In regard to the biocide specific study
380 design elements, the same principles apply as for the total residue studies.

381 **5.2.3. Other uses of the substance**

382 While the European Medicines Agency is only responsible for detailed evaluation of consumer exposure
383 to biocidal substances used in animal husbandry, it is appropriate that any consumer exposure to the
384 substance that may occur as a result of other uses of the substance should be taken into account. The
385 dossier submitted by a company seeking authorisation of a product should therefore include
386 information on all known uses of the pharmacologically active substance along with a calculation of the
387 proportion of the ADI used as a result of consumer exposure to residues resulting from uses of the
388 pharmacologically active substance in products other than biocidal products for use in animal
389 husbandry.

390 **6. Derivation of the MRL**

391 The general principles underlying the derivation of numerical MRLs are set out in Volume 8. However,
392 as described in sections 4.1.2.1 and 4.1.2.2 there may be specific cases in which MRLs can be derived
393 based on limited data packages. The establishment of numerical MRL values will always require
394 availability of a validated analytical method for residue surveillance, as described in Volume 8.

395 **7. Extrapolation of MRLs (including 'no MRL required' status)**

396 Volume 8 also sets out principles by which MRLs may be extrapolated within groups of species and,
397 where MRLs have been established for a major ruminant species, a major monogastric species, and for
398 chickens, to all food producing species¹¹.

¹¹ For further information on the definition of major and minor species see the CVMP Position paper regarding availability of products for minor uses and minor species (MUMS) (EMA/CVMP/477/03/Final).

399 Definitions

400 **Exposure reduction measure:** A restriction to the way in which a product is used that has the effect
401 of reducing the exposure of consumers to residues of the pharmacologically active substance.
402 Examples of exposure reduction measures include withdrawal periods, removal of animals from the
403 application environment during product application, rinsing walls/equipment after product application.
404 Exposure reduction measures incorporated into product literature should be demonstrated to lead to
405 residue levels that conform to established MRLs.

406 **External exposure:** Exposure reaching the outside the animal's body boundary (for example, on the
407 skin, in lungs, in the gastro-intestinal tract). External exposure is not adjusted for factors such as
408 dermal absorption, oral absorption or breakdown in the digestive system of the livestock animal or
409 absorption via the livestock animal's inhalatory system.

410 **Internal exposure:** (Systemic) exposure of the body after passage of the body boundaries. Internal
411 exposure is the bioavailable fraction of the external exposure, which determines the amount of
412 residues in the target tissues of food producing animals.

413 References

414 Volume 8 of The rules governing medicinal products in the European Union: Notice to applicants and
415 guideline. Establishment of maximum residue limits (MRLs) for residues of veterinary medicinal
416 products in foodstuffs of animal origin (2005). Available at
417 http://ec.europa.eu/enterprise/sectors/pharmaceuticals/documents/eudralex/vol-8/index_en.htm

418 CVMP Position paper regarding availability of products for minor uses and minor species (MUMS)
419 (EMA/CVMP/477/03/Final).

420 European Commission Draft Guidance on Estimating Livestock Exposure to Active Substances used in
421 Biocidal Products. To be published at <http://ec.europa.eu/environment/biocides/consultation.htm>

422 OECD 505 "Residues in Livestock" for guidance on duration of feeding studies "*Once acclimatized,*
423 *animals should be dosed daily for a minimum of 28 days or until residues plateau in milk or eggs, if*
424 *they have not done so in 28 days"*
425 [http://lysander.sourceoecd.org/vl=2617880/cl=59/nw=1/rpsv/cw/vhosts/oecdjournals/1607310x/v1n](http://lysander.sourceoecd.org/vl=2617880/cl=59/nw=1/rpsv/cw/vhosts/oecdjournals/1607310x/v1n7/contp1-1)
426 [7/contp1-1](http://lysander.sourceoecd.org/vl=2617880/cl=59/nw=1/rpsv/cw/vhosts/oecdjournals/1607310x/v1n7/contp1-1)

427 Leeman et al. (2007): Transfer of chemicals from feed to animal products: The use of transfer factors
428 in risk assessment. Food additives and contaminants; 24,1-13

429 **Annex I - Derivation of the threshold value of 4 µg/kg**
 430 **bw/day for external exposure of food producing animals**

431 The threshold value of 4 µg/kg bw/day, summed over all exposure routes, for the external exposure of
 432 an animal was established by the Biocides Technical Meeting at its meeting of 16-20 March 2009. The
 433 trigger value is extrapolated from a value used by the European Food Safety Authority (EFSA) in its
 434 assessments of plant protection products under Directive 91/414/EC. EFSA decides whether to initiate
 435 the process of food risk assessment and possible MRL setting in food of animal origin based on the
 436 substance content of the animal feed, which in turn determines the animal's exposure to the
 437 substance. The threshold value used by EFSA is 0.1 mg of substance per kg of feed dry matter. The
 438 EFSA trigger value for substance content in animal feed was extrapolated to a value for the external
 439 dose of a biocidal substance using standard livestock weights and feed intake.

440 The data on animal weights and feed intake were taken from Appendix G of the DG SANCO Guidelines
 441 for the generation of data concerning residues as provided in Annex II part A, section 6 and Annex III,
 442 part A, section 8 of Directive 91/414/EEC concerning the placing of plant protection products on the
 443 market (<http://ec.europa.eu/food/plant/protection/resources/app-g.pdf>, which is available at
 444 http://ec.europa.eu/food/plant/protection/resources/publications_en.htm#residues).

445 The results of the calculations are shown in the following table:

446

	Chicken	Dairy cattle	Beef cattle	Pig	Model Goat	UK Sheep	UK Turkey
Body weight [kg] - default	1.9	550	350	75	70	75	7
Feed (dry matter) intake [kg /day] - default	0.12	20	15	3	3	3	0.2
Substance intake [mg/day] at the 0.1 mg/kg feed trigger value	0.012	2	1.5	0.3	0.3	0.3	0.02
Substance intake [mg/kg bw/ day]	0.0063	0.0036	0.0043	0.0040	0.0043	0.0040	0.0029

447

448 The first 4 columns of the above table correspond to the 4 indicator livestock species described in the
 449 SANCO guidance (chicken including laying hens, dairy cattle, beef cattle, pig). The additional 3
 450 columns (Model goat, UK sheep and UK turkey) provide values commonly accepted within EFSA.

451 As expected, the resulting substance intake values differ between species. However, because the
 452 variation range is narrow, because the value of 0.1 mg/kg feed dry matter is already considered
 453 conservative, and because there is no need for absolute precision for an indicator of need for further
 454 refinement, it is considered that the median value of 0.004 mg/kg bw (4 µg/kg bw) for external
 455 exposure over 1 day can be accepted as a threshold value that provides similar level of conservatism
 456 to the trigger value used by EFSA in its evaluation of plant protection products. The trigger value of
 457 4 µg/kg bw/day is considered appropriate for use in relation to all livestock species

458 **Annex II – Defining substances of particular concern**

459 EFSA uses a trigger value of 0.1 mg of active substance per kg feed, in order to determine whether the
 460 establishment of MRLs in food of animal origin needs to be considered for a plant protection product.
 461 As described in Annex 1, this was calculated to correspond to an external dose of 4 µg/kg bw/day
 462 received by the animal.

463 By using a conservative human exposure estimate it can be calculated that a level of 0.1 mg/kg feed
 464 would lead to an estimated human intake of 293 µg per person per day. The calculation assumes oral
 465 exposure of food producing animals to 0.1 mg/kg feed, and uses transfer factors (Leeman et al, 2007)
 466 to estimate the amount of substance transferred from animal feed to food commodities, and the CVMP
 467 food basket to calculate the theoretical maximum daily intake of the substance. The transfer factors
 468 used in the calculation were the values established for the most conservative class of compounds, i.e.
 469 compounds with a Log P_{o/w} between 6 and 7. It is assumed that the biocides to be evaluated are within
 470 the domain of the chemicals assessed in the above mentioned study. The resulting theoretical
 471 maximum daily intake for the foodstuffs is shown in the table below.

	P₉₅ for the transfer factor	Estimated content in commodity after oral exposure of 0.1 mg/kg feed (µg/kg)	Food basket (kg)(calculation of maximum theoretical daily intake for consumers)	Estimated maximum theoretical daily intake for humans using the food basket (µg)
Egg	1,60	160,00	0,10	16,00
Milk	0,52	52,00	1,50	78,00
Meat	0,33	33,00	0,30	9,90
Fat	30,00	3000,00	0,05	150,00
Liver	2,62	262,00	0,10	26,20
Kidney	2,62	262,00	0,05	13,10
				293,20

472

473 Thus the EFSA trigger value of 0.1 mg of active substance per kg feed is anticipated to lead to a TMDI
 474 of 293 µg per person. Therefore, if the ADI of the substance under examination is above 293 µg per
 475 person (or 5 µg/kg bw), it can be concluded that the external exposure of the animals at the
 476 established trigger value will lead to a TMDI below the ADI and so consumer safety will be ensured.
 477 This assumption is made using very conservative transfer factors and a very conservative human
 478 exposure scenario.

479 To obtain an idea of how protective the trigger value is, this ADI of 5 µg/kg bw was correlated with the
 480 ADIs of some known potent pesticide substances. The vast majority of these substances have ADIs
 481 well above the cut-off value, and would therefore not represent a risk for the consumer at the
 482 threshold value of 4 µg/kg bw/day.

483 However, a number of pesticides have ADIs below the cut-off value. It seems to be especially
 484 cholinesterase inhibitors (neurotoxicants) and some substances with effects on liver and/or kidneys,
 485 and there is one example of a substance causing anemia. A number of these substances are classified
 486 as reproductive toxicants.

487 It can be concluded that while the trigger value approach can be safely applied in a majority of cases,
 488 it should not be used for substances of particular concern, i.e. those with a potential for nonthreshold
 489 effects (for example, genotoxic effects), or for reproductive/developmental/neurotoxic actions or other

490 critical endpoints. Some of these substances potentiate their action because they accumulate in the
491 organism, so this physico-chemical property should also be included in the identification of substances
492 of particular concern. In general, substances with a log Pow greater than 3 can be considered to have
493 the potential to accumulate. Any biocide for which it is estimated that the ADI will be below 5 µg/kg bw
494 or for which there is the suspicion of non-threshold effects or toxicity at low doses, may present a
495 possible risk for the consumer and should therefore lead to the triggering of a request for MRL
496 assessment even though the external exposure of the animal may be below the threshold value of
497 4 µg/kg bw/day.